Exposure to fine and ultrafine particles from secondhand smoke in public places before and after the smoking ban, Italy 2005

Pasquale Valente, Francesco Forastiere, Antonella Bacosi, Giorgio Cattani, Simonetta Di Carlo, Monica Ferri, Irene Figà-Talamanca, Achille Marconi, Luigi Paoletti, Carlo Perucci, Piergiorgio Zuccaro

**Background:** A smoking ban in all indoor public places was enforced in Italy on 10 January 2005.

**Methods:** We compared indoor air quality before and after the smoking ban by monitoring the indoor concentrations of fine (≤2.5 μm diameter, PM$_{2.5}$) and ultrafine particulate matter (<0.1 μm diameter, UFP). PM$_{2.5}$ and ultrafine particles were measured in 40 public places (14 bars, six fast food restaurants, eight restaurants, six game rooms, six pubs) in Rome, before and after the introduction of the law banning smoking (after 3 and 12 months). Measurements were taken using real time particle monitors (DustTRAK Mod. 8520 TSI; Ultra-fine Particles Counter-TRAK Model 8525 TSI). The PM$_{2.5}$ data were scaled using a correction equation derived from a comparison with the reference method (gravimetric measurement). The study was completed by measuring urinary cotinine, and pre-law and post-law enforcement among non-smoking employees at these establishments.

**Results:** In the post-law period, PM$_{2.5}$ decreased significantly from a mean concentration of 119.3 μg/m$^3$ to 38.2 μg/m$^3$ after 3 months (p < 0.005), and then to 43.3 μg/m$^3$ a year later (p < 0.01). The UFP concentrations also decreased significantly from 76 956 particles/cm$^3$ to 38 079 particles/cm$^3$ (p < 0.0001) and then to 51 692 particles/cm$^3$ (p < 0.01). Similarly, the concentration of urinary cotinine among non-smoking workers decreased from 17.8 ng/ml to 5.5 ng/ml (p < 0.0001) and then to 3.7 ng/ml (p < 0.0001).

**Conclusion:** The application of the smoking ban led to a considerable reduction in the exposure to indoor fine and ultrafine particles in hospitality venues, confirmed by a contemporaneous reduction of urinary cotinine.

The application of the law banning smoking in indoor public places in Italy on 10 January 2005 follows other smoking bans in the world. The Smoking Ordinance (Labor code 6404.5) of January 1998, extended the smoking ban to bars, game rooms, bingo parlours, casinos, and other public locations in California, and made it the first state to outlaw smoking in all indoor public places. Similar legislation has been subsequently adopted in India, Ireland, Malta, New Zealand (2004), Norway, and South Africa (2000). It is expected that smoking bans will be adopted in several other European countries in the near future, although the introduction of laws that prohibit smoking have often been highly contested.

Environmental tobacco smoke (ETS) is an air pollutant made up of a complex mixture of around 4000 chemicals, some 50 of which are carcinogens. Exposure to ETS is causally associated with several adverse health effects, including lung cancer and heart disease in adults, exacerbation of asthma, lower respiratory illnesses, ear infections, and other diseases in children and infants. A recent report of the European Respiratory Society estimates that a total of 80 000 adults in Europe in 2002 died from exposure to ETS, including 7000 in Italy. The annual average guideline value of 10 μg/m$^3$ for PM$_{2.5}$ was recently issued a global update of the Air Quality Guidelines. The average annual guideline value of 10 μg/m$^3$ for PM$_{2.5}$ was chosen to represent the lower end of the range over which

Vapour phase nicotine and particulate matter (PM$_{2.5}$) are the most commonly used indicators to evaluate environmental exposure to ETS. Numerous studies have shown that tobacco smoke plays a major part in determining the concentration of particulate matter indoors. In addition, ETS is, like all other byproducts of combustion, made up of a considerable number of ultrafine particles (UFP), with a diameter <0.1 μm, since the average aerodynamic diameter of ETS particles is approximately 0.2 μm. Ultrafine particles are an important cause of adverse effects as a consequence of their large total surface, their large alveolar deposition, inflammatory capability, and their possibility to translocate in the general circulation. Main indoor sources of UFP are smoking, cooking, candles, and chemical reactions (for example, terpenes and ozone). Nicotine is principally metabolised into cotinine by the liver; it has a half life of around 18 hours and is generally considered the best indicator of nicotine intake. Urinary cotinine levels correlate with levels of environmental nicotine and with self reported ETS exposure.

The evidence of the health effects of airborne particles comes from several lines of scientific investigation, from cellular and animal studies to large epidemiological investigations. Health effects include premature mortality, cardiorespiratory morbidity, and lung cancer. The World Health Organization has recently issued a global update of the Air Quality Guidelines. The average annual guideline value of 10 μg/m$^3$ for PM$_{2.5}$ was chosen to represent the lower end of the range over which
significant effects on survival have been observed. The 24 hour PM$_{2.5}$ guideline is 25 µg/m$^3$. While there is considerable toxicological evidence of potential detrimental effects of ultrafine particles on human health, the existing body of epidemiological evidence was insufficient to reach a conclusion on the exposure–response relation to ultrafine particles and no recommendation was provided in the guideline for concentrations of ultrafine particles.

In order to evaluate the efficacy of the new legislation in public indoor environments in Italy, we studied changes in environmental concentrations of fine and ultrafine particles, as well as urinary cotinine in employees of public establishments, before and after (in the third and 12th month) the implementation of the law banning smoking.

MATERIALS AND METHODS
We measured exposure to environmental tobacco smoke by determining PM$_{2.5}$ and the number of ultrafine particles in 40 establishments in the city of Rome (14 bars, six fast food restaurants, eight restaurants, six video game parlours, six pubs). The 40 locations were selected randomly from the official list of businesses in the western part of the city (Health District D, about 400 000 inhabitants). The venues’ owners were contacted by the regional health authority to request participation in the study. Baseline measurements of the environment were taken in November/December 2004 (before the law was in effect). In March/April 2005, and November/December 2005 (after the law was in effect) the comparison measurements were taken again, without warning and during peak business hours. To increase the statistical efficiency and avoid a “learning effect,” the number of locations studied was constant (40), but 50% of them were rotated out of the study for each successive measurement, and replaced with other establishments.

PM$_{2.5}$ was measured with a real time aerosol monitor (DustTrak, Model 8520 TSI), equipped with a sampling nozzle to collect fine particles (PM$_{2.5}$). In the same establishments, ambient ultrafine particles were counted using a condensation particle counter: P-Trak Monitor (Mod. 8525 UPC TSI). At the same time, continuous measurements were taken of temperature, relative humidity, CO$_2$, and air velocity (with Q-Trak Plus TSI). A survey of all these parameters was taken in the outside environment for purposes of comparison. The instruments took measurements every 10 seconds, and recorded the average reading every minute, in both smoking and non-smoking areas. Every day the instruments were recalibrated (to zero). Instruments were positioned inside, an average height of 0.8 metre off the ground, and weather and traffic conditions were recorded.

PM$_{2.5}$ measurements with DustTrak were later gravimetrically calibrated based on empirically derived correction equations. For one month in winter and one month in summer, in two environments, one reserved for smokers and the other for non-smokers, PM$_{2.5}$ measurements were taken both with DustTrak and an active air pump (Skypost PM TCR-Tecora) with an air flow of 2.3 m$^3$/h conforming to specified European protocols for measuring PM$_{2.5}$.$^{14}$ The particles were collected on a 47 mm filter membrane. Each measurement was conducted over 8 hours. The measurements were carried out daily, in two phases (first phase from 8.00–16.00, the second phase from 20.00–04.00). The mass concentration of particulate collected was determined by manual gravimetric method, using a Sartorius scale with a resolution of 1 µg, after the filters were pretreated for at least 48 hours at 50 (SD 5) RH%, 293 (SD 1) K (in a room with controlled temperature and relative humidity). Based on the linear regression of PM$_{2.5}$ concentrations measured with Dust Trak/manual gravimetric method (RFM), two correction equations were applied to the PM$_{2.5}$ data: $x = (y+21.01)/4.01$ for smoking environments ($y = 4.0132 (1.0544) x - 21.048 (72.5159)$) $R^2 = 0.7567$, and $x = (y+9.1)/2.66$ for smoke free environments ($y = 2.6587 (0.1932) x - 9.108 (3.8360)$) $R^2 = 0.917$.

All non-smoking employees of the 40 facilities were asked to provide a urine sample, contemporaneously with PM$_{2.5}$ measurements. An interview was completed for each subject, on personal data and information to evaluate subjective exposure to passive smoke in the workplace and at home (presence of smoke, hours of exposure per week). Urine samples were stored at −80°C. Urinary cotinine was measured using radioimmunoassay (RIA), according to the method described by Van Vuurink et al.$^{26}$ A cut-off of 100 ng/ml was employed to distinguish active smokers from non-smokers. Two subjects of the first phase were excluded from the analyses with cotinine levels of >100 ng/ml.

Arithmetic and geometric means were calculated. Levels of environmental and biological data from the three periods were compared with the Mann-Whitney test. The results were confirmed using separate linear regression analysis with the log transformation of PM$_{2.5}$, UF, and urinary cotinine as the dependent variable and an indicator variable for the three periods of the assessment as the independent variable. The data were analysed using Stata 8 software.

RESULTS
Table 1 reports concentration levels of PM$_{2.5}$ measured indoors before and after the law (in the third and 12th month); values recorded outdoors are also reported. The indoor microclimatic parameters did not show significant differences over the three measurement periods, and the average temperature remained steady (21°C before, 20.8°C first post-law period, and 20.9°C second post-law period). In 40 establishments monitored before the law took effect, 17 (42.5%) were completely non-smoking, 14 (35%) allowed smoking anywhere, and nine (22.5%) reserved separate smoking areas. The prohibition nonetheless was not enforced in all of the video game parlours. The average PM$_{2.5}$ value for all smoking areas before the smoking ban was 184.6 µg/m$^3$ (95% CI 104 to 264.8). The highest values were observed in pubs that allowed smoking (mean 368.1 µg/m$^3$, 95% CI 89.3 to 646.9). In the areas reserved for non-smokers before the ban, readings of PM$_{2.5}$ showed a mean of 56.7 µg/m$^3$ (95% CI 34.9 to 78.5), higher than the values found in the air measured directly outside (mean 24.6, 95% CI 19.9 to 29.4). Statistical significant reductions were found in average PM$_{2.5}$ values: from a pre-law level of 119.3 µg/m$^3$, which fell to 38.2 µg/m$^3$ and 43.3 µg/m$^3$ after the ban took effect. A reduction in particles was observed even in businesses that had never allowed smoking. In restaurants, video game parlours and pubs, the reduction of PM$_{2.5}$ concentrations was particularly strong. Outdoor PM$_{2.5}$ concentrations were relatively stable and close to the values recorded at a fixed monitoring site located at the National Health Institute, in the centre of the city.

A reduction of UF values was also observed, although not as notable as for fine particles. The particle number concentration
went from 76,956 pt/cm³ to 38,079 pt/cm³ ($p < 0.0001$) and then 51,692 ($p < 0.01$) (table 2). It should be noted that outdoor concentrations of ultrafine particles were lower in the first post-law period than in the pre-law survey but the concentrations were higher in the second post-law assessment, again following the pattern monitored at the fixed station.²¹

Thirty-nine subjects agreed to provide urine samples in the pre-law monitoring period (two were excluded), 23 in the third...
month post-law, and 11 in the 12th month post-law. The median age in the first period was 39 years, and 34 years in the post-law periods. The reduction in exposure observed in the biological readings (table 3) appears to be consistent with those found in the environmental measurements. Before the ban, the 37 subjects who were analysed (two excluded as smokers) presented a mean of urinary cotinine concentration of 17.8 ng/ml (95% CI 14 to 21.6, SD 11.4, median 14.2). Three months after the ban, in the 23 subjects examined, the mean urinary cotinine concentration fell to 5.5 ng/ml (95% CI 3.8 to 7.2; SD 4.3, median 6.0). A year later there was a striking reduction in urinary cotinine (mean 3.7, 95% CI 1.8 to 5.6; SD 2.8, median 4.0). The reduction was particularly notable in bars and restaurants (fig 1). Also, the analysis of the questionnaires showed a significant reduction of subjective exposure to ETS at the workplace in the post-law periods (p<0.0005).

When we considered the data from the three surveys, there was a statistically significant association between both PM$_{2.5}$ and UFP levels and urinary cotinine concentrations. In particular, there was an increase of 0.75 ng/ml (95% CI 0.4 to 1.1) in urinary cotinine for each 10 $\mu$g/m$^3$ increase in PM$_{2.5}$. The association between PM$_{2.5}$ and urinary cotinine is shown in figure 2.

### DISCUSSION

We found that the application of the smoking ban led to a considerable reduction in exposure to indoor fine and ultrafine particles in public facilities. The results were confirmed with a simultaneous reduction of urinary cotinine.

Few studies have reported objective ETS exposure data from public locations, even fewer have used environmental PM$_{2.5}$ and cotinine concentrations as the exposure indicator. Measurements taken in residences and offices indicate good correlation between ambient nicotine and ambient particulate concentrations. This has been particularly true in locations studied where (a) there is regular smoking, (b) microenvironments are in a steady state, and (c) the measurement periods

### Table 3 Urinary cotinine values in employees of public establishments, before and after the smoking ban

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<th>Men</th>
<th>Women</th>
<th>Total</th>
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are significantly longer than necessary for smoke diffusion. PM$_{2.5}$ is a widely accepted indicator to evaluate indoor air quality and is a valid exposure marker of ETS.

The real time measurements method (DustTrak) was well correlated with the gravimetric reference method, but it constantly overestimated PM$_{2.5}$ levels (2.66 times higher), and the overestimation was even higher in the presence of ETS (4.01). The values measured with DustTrak, once corrected, resulted in line with measurements from other studies that used the gravimetric method. The mean (corrected) PM$_{2.5}$ from the entire pre-law sample was 119.3 g/m$^3$. Recently, measurements with the gravimetric method showed average concentrations of PM$_{2.5}$ to be 114.5 g/m$^3$ in the smoking sections of 59 pubs and bars. Siegel reported a weighted respirable suspended particle (RSP) average of 117 g/m$^3$, based on 12 studies in restaurants. These values are analogous to those reported by Ellingsen$^{24}$ for total particulate in restaurants in Norway (115 µg/m$^3$). Outdoor concentration levels after correction also appear to be analogous to those calculated with the reference method for Rome.

Our data confirm the presence of elevated levels of ETS exposure in indoor hospitality venues, similar to those observed in England,$^{28}$ in Ireland,$^{26}$ in Norway,$^{25}$ and in the US,$^{29-31}$ before the smoking bans in restaurants and bars. Those who work in environments where there are no smoking restrictions for customers have three times higher exposure levels to ETS than employees of businesses where smoking is restricted to a special area. They also present cotinine levels that are much higher than employees of completely non-smoking businesses.

As in Ireland, Norway, and the US, smoking bans in Italy also have drastically reduced ETS pollution. Our study showed that PM$_{2.5}$ was reduced by two-thirds and urinary cotinine by 73%. A statistically significant reduction in ETS pollution was also seen in establishments that already prohibited smoking. The significant reduction in air pollution is probably the result of both the rigorous manner in which the law was enforced, and of the great reduction in the number of businesses with smoking sections. In smoking sections, the reduction in the particulate concentrations was markedly smaller. One study that measured the changes in RSP levels in 20 restaurants and bars in New York reported an 84% decrease after the Clean Indoor Air Act (CIAA) was put into effect in July 2003.$^{32}$ Similar results were obtained by Repace in Delaware. Examining the changes produced by the CIAA on the air quality, he found that RSP decreased by 91%, contemporaneously to a 95% reduction of polycyclic aromatic hydrocarbons.$^{33}$ In a cross sectional study of the hospitality industry, after the CIAA took effect in New York, a significant reduction was shown in urinary cotinine and ETS exposure levels, compared to pre-law levels.$^{34}$ In particular, the proportion of subjects with undetectable levels of urinary cotinine increased from 3% to 62% after the law, while average values of urinary cotinine decreased from 4.93 ng/ml to 0.30 ng/ml. These studies indicate that most (70–90%) fine particulate concentrations indoors are related to cigarette smoke and that eliminating smoke significantly improves the quality of indoor air, with substantial reduction of health risks. Reductions similar to ours have been verified by using other markers, like saliva cotinine (reduced by 70%) by Mulcahy et al.$^{35}$ in Ireland. In that study, self reported exposure to passive smoke showed a significant reduction from an average of 30 hours to zero hours and was associated with a great reduction (83%) in concentrations of nicotine in the air, from an average of 35.5 mg/m$^3$ to 5.95 mg/m$^3$. Ellingsen described exposure to nicotine and indoor particles, before and after the smoking ban was extended to include bars and restaurants in Norway. The mean concentrations of nicotine and total particulate (with gravimetric method) decreased from 28.3 µg/m$^3$ (range 0.4–88.0) and 262 µg/m$^3$ (range 52–662), respectively, to 0.6 µg/m$^3$ (range 0–3.7) and 77 µg/m$^3$ (range 0–261) after the ban. Pearson’s correlation between nicotine in the air and total particulate was very good (0.86, p<0.001; n = 48). The geometric average of urinary cotinine concentrations was reduced from 9.5 µg/g creatinine (95% CI 6.5 to 13.7) to 1.4 µg/g creatinine (95% CI 0.8 to 2.5) (p<0.001) in 25 nonsmokers.$^{36}$ Similarly to Bates,$^{37}$ our study also shows that employees of hospitality venues with smoking sections reserved for their customers are still considerably more exposed than those who work in non-smoking businesses.

Our data also suggest a relation between ETS pollution and particulate matter. Smoking is an important source of indoor UFP, but their number is also dependent on other sources like cooking appliances and burning candles; this could explain the lower relative decrease of UFP when compared to PM$_{2.5}$ and urinary cotinine.

**CONCLUSION**

The introduction of the smoking ban in the workplace is a key tool for public health administrators determined to reduce passive smoke exposure in the general population. The examples available include Ireland, New York, New Zealand, Norway and now Italy. Such successes demonstrate that it is possible to significantly reduce ETS exposure. As a result, the incidence of smoking related diseases is bound to decrease, but the full effects have yet to be studied.$^{38,39}$ An 11% reduction in hospitalisation rates for myocardial infarction among people under 60 years of age has been noted in one Italian region in the six month period following the smoking ban.$^{40}$ Exposure reduction is predominantly associated with the elimination of smoking areas, while in smoking sections, in spite of air exchange and recycling systems, significant differences in air quality exist. The results of environmental and biological monitoring indicate that in facilities with smoking areas, a residual risk exists for the health of the employees.

**ACKNOWLEDGEMENTS**

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**What is already known on this subject**

Some studies have suggested a positive impact of smoking bans on the reduction of ETS exposure. The studies have used airborne nicotine or particulate matter and urinary cotinine as exposure markers.
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